

1 **Arsenic accumulation and speciation in strawberry plants exposed to inorganic arsenic**  
2 **enriched irrigation**

3 Ana Isabel González de las Torres<sup>1</sup>, Inmaculada Giráldez<sup>2</sup>, Fátima Martínez<sup>3</sup>, Pedro Palencia<sup>4</sup>,  
4 Warren T. Corns<sup>5</sup>, Daniel Sánchez-Rodas<sup>1,2\*</sup>

5 <sup>1</sup> Center for Research in Sustainable Chemistry-CIQSO. University of Huelva. 21071-Huelva, Spain.

6 <sup>2</sup> Department of Chemistry. Faculty of Experimental Sciences. University of Huelva. 21071-Huelva.

7 <sup>3</sup> Department of Agroforestry Sciences. ETSI. University of Huelva. 21071-Huelva, Spain.

8 <sup>4</sup> Department of Organisms and Systems. Polytechnic School of Mieres. University of Oviedo. 33600-  
9 Mieres, Spain.

10 <sup>5</sup> P S Analytical Ltd., Arthur House, Crayfields Industrial Park, Main Road, Orpington, Kent, UK.

11

12 **ABSTRACT**

13 The accumulation and transformation of arsenic species have been studied in the context of  
14 hydroponic cultivation of strawberry plants. Cultivation experiments have been performed by  
15 adding inorganic arsenic at concentrations of 10, 100 and 1000  $\mu\text{g L}^{-1}$  via root irrigation. The  
16 total arsenic content was determined by Hydride Generation-Atomic Fluorescence  
17 Spectrometry (HG-AFS). The accumulation was dependent on the concentration of arsenic  
18 added to the irrigation and the arsenic species. Arsenic (III) accumulated at higher rates than  
19 arsenic (V). A greater accumulation of arsenic was found in roots (0.44-4.10  $\text{mg kg}^{-1}$ ) than in  
20 stems (0.43-1.27  $\text{mg kg}^{-1}$ ) and fruits (0.22-0.30  $\text{mg kg}^{-1}$ ). The speciation results obtained by  
21 HPLC-HG-AFS analysis indicated that the addition of As(III) resulted in a partial methylation  
22 producing monomethyl arsenic (MMA) and dimethyl arsenic (DMA). After As(V) addition,  
23 only MMA was observed and this was accompanied with a notable reduction in the ratio of  
24 As(V) to As(III).

25

26 **Keywords:** strawberry plants; irrigation; arsenic; speciation; fruits; HPLC-HG-AFS

## 27 **1. Introduction**

28 Arsenic is regarded as a toxic and carcinogenic element (IARC, 2004). High  
29 concentrations of arsenic in drinking water, soil, crops, vegetables, animals and the food chain  
30 may cause human health problems (Nagajyoti, Lee, & Sreekanth, 2010; Zhao, McGrath &  
31 Meharg, 2010; Abdul, Jayasinghe, Chandana, Jayasumana, & De Silva, 2015). Following the  
32 overwhelming accumulation of evidence relating to the chronic toxicological effects of arsenic,  
33 the World Health Organization guideline now recommends a maximum of  $10 \mu\text{g L}^{-1}$  of arsenic  
34 in drinking water (WHO, 2011a). Some legal regulations for arsenic in animal feed and food  
35 have also been introduced, and they are focused on inorganic arsenic defined as the sum of  
36 arsenite (As(III)) and arsenate (As(V)) (Petursdottir, Sloth, & Feldmann, 2015).

37  
38 Food safety is of paramount public concern, as relatively high amounts of arsenic may be  
39 present in some foods (Sirot, Guérin, Volatier, & Leblanc, 2009). Elevated arsenic  
40 concentrations are naturally found in seafood (Taylor et al., 2017). This seems to be related to  
41 the relatively high concentration of arsenic ( $1.5\text{-}2 \mu\text{g L}^{-1}$ ) in seawater, which is then ingested  
42 by marine life (Urgast, Adams, Raab, & Feldmann, 2010). In contrast, foods of terrestrial origin  
43 such as cereals, meats, vegetables, fruits and dairy products usually contain a lower arsenic  
44 concentration ( $<0.1 \text{ mg kg}^{-1}$ ) (Sigrist, Hilbe, Brusa, Campagnoli, & Beldomenico, 2016). In  
45 some cases, arsenic concentrations of up to  $1.27 \text{ mg kg}^{-1}$  and  $2.20 \text{ mg kg}^{-1}$  have been reported  
46 for fruits and vegetables, respectively (WHO, 2011b). The presence of arsenic in crops,  
47 vegetables, animals and food products has been recently reviewed (Upadhyay, Shukla, Yadav,  
48 & Srivastava, 2019). Since inorganic arsenic is the most toxic fraction of this metalloid, the  
49 Food and Agriculture Organization (FAO) and WHO initially recommended a provisional  
50 tolerable weekly intake (PTWI) of  $15 \mu\text{g}$  of inorganic As  $\text{kg}^{-1}$  body weight (WHO, 2011a).  
51 However, in 2010 this value was no longer considered appropriate. Therefore, a benchmark

52 dose (BMDL<sub>0.5</sub>) was introduced and it noted a 0.5% increased incidence of lung cancer with an  
53 exposure of 3.0 µg kg<sup>-1</sup> body weight per day (WHO, 2011b). The inorganic arsenic fraction  
54 represents 28-100% of the total arsenic concentration in the edible tissues of vegetables (carrot,  
55 garlic, potato, and beetroot), depending on the species, growth stage and organ (Munoz et al.,  
56 2002).

57

58 Hence, there is a need for robust validated methods for selective extraction and  
59 determination of inorganic arsenic in food matrices, as indicated by international agencies  
60 (WHO, 2011b). In order to refine risk assessment of inorganic arsenic, it is also required to  
61 produce data regarding the different arsenic species in food commodities to support dietary  
62 exposure assessment and dose-response data for the possible health effects (EFSA, 2009). In  
63 this sense, there is a research gap relating to the development of low cost and simple analytical  
64 methods and techniques to reliably determine arsenic species and inorganic arsenic in foods.

65

66 The topic of elemental speciation is a well-established area of research. Studying the  
67 chemical forms of the elements helps to elucidate their mobility, biological availability,  
68 distribution and toxicity (Ure & Davidson, 2002; Gupta, Nayak, Agarwal, Dobhal, Uniyal, &  
69 Singh, 2012). Inorganic arsenic compounds are generally considered more toxic than the  
70 organic compounds (Mattusch, Wennrich, Schmidt, Reisser, & Fresenius, 2000). The toxicity  
71 of arsenic species follows the order As(III) > As(V) > MMA (monomethyl-arsenic) > DMA  
72 (dimethyl-arsenic). There are reports of transformation of arsenic species in plant systems  
73 (Kuehnelt, Lintschinger, & Goessler, 2000). Therefore, the quantification of arsenic species in  
74 vegetables is of great importance are important for investigation of their accumulation and  
75 metabolism in the leaves, roots, shoots, grains or fruits of edible plants grown in different  
76 conditions (Zhao, McGrath, & Meharg, 2010).

77 Speciation analysis in vegetables is achieved by extraction of arsenic species followed by  
78 analytical techniques that involve liquid chromatography coupled to atomic spectroscopy  
79 and/or mass spectrometry (Zhao, Li, Xu, Luo, & Ma, 2015; Sons, Lee, Kim, Lee, & Nam,  
80 2019). As an alternative to liquid chromatography, total inorganic arsenic can be determined  
81 by selective arsine generation with atomic absorption or atomic fluorescence spectrometry as  
82 under the conditions of high acidity DMA does not form a hydride. This approach however  
83 relies on the criteria that no MMA is present in the sample which is often the case for most food  
84 samples. (Chen, Corns, Stockwell & Huang, 2014; Cerveira, Pozebon, Poméu de Moraes, &  
85 Silva de Fraga, 2015).

86

87 Most of the studies on arsenic speciation in food plants are focused on rice and its  
88 products (flour), which is due to its importance as a major food. Rice cultivation is sometimes  
89 affected by arsenic polluted irrigation (Abedin, Cresser, Meharg, Feldmann, & Cotter-Howells,  
90 2002; Farías, Londonio, Quintero, Befani, Soro, & Smichowski, 2015; Dos Santos, Pozebon,  
91 Cerveira, & de Moraes, 2017). Arsenic speciation dynamics in rice cultivation have been  
92 recently reviewed, highlighting the reduction of As(V) to highly mobile As(III) during flooding,  
93 as well as the roles of microorganisms in methylation reactions (Kumarathilaka, Seneweera,  
94 Meharg, & Bundschuh, 2018). Other horticultural vegetables studied in relation to arsenic  
95 exposure are mustard (Jedynak, & Kowalska, 2011), bean (Sadee, Foulkes, & Hill, 2016), carrot  
96 (Mayorga, Moyano, Anawar, & García-Sánchez, 2013; Bergqvist, Herbert, Persson, & Greger,  
97 2014), spinach (Shahid et al., 2017) and lettuce (Ma, Yang, Kong, & Wang, 2017).

98

99 There is scarce information available regarding the accumulation of arsenic in strawberry  
100 plants (*Fragaria ananassa* Duch), a cultivation with worldwide importance. Strawberry  
101 cultivation is of particular relevance in Spain, the leading strawberry producer in Europe.

102 Approximately 90% of the Spanish production is concentrated in the province of Huelva (SW  
103 Mainland Spain), which occupies a surface of ca. 7000 Ha with an estimated annual production  
104 of  $3.6 \times 10^5$  tonnes in 2016 (MAPAMA, 2011; FAO, 2018). The presence of arsenic in some  
105 areas of the aquifer near the farming areas (Kohfahl, Sánchez-Rodas Navarro, Mendoza,  
106 Vadillo, & Giménez-Forcada, 2016) motivates the study of the response of strawberry plants to  
107 possible arsenic exposure. Also, there is no information available about the arsenic uptake rate  
108 into different parts of the strawberry plants, as well as a lack of knowledge on species  
109 transformation within the fruits.

110  
111 The purpose of this work is to study the accumulation of arsenic in strawberry plants  
112 irrigated with arsenic enriched water (arsenite and arsenate), with consideration of roots, stems  
113 and fruits. The possible transformation of inorganic arsenic into less toxic methylated species  
114 (MMA, DMA) in the fruits has been studied. Speciation analysis of arsenic with liquid  
115 chromatography coupled to atomic fluorescence spectrometry has been performed. This  
116 information will allow us to know whether the inorganic arsenic accumulated in the fruits  
117 represents a health risk.

118

## 119 **2. Materials and methods**

### 120 *2.1. Reagents*

121 Standard solutions of  $1000 \text{ mg L}^{-1}$  of As(III), DMA, MMA and As(V) species were prepared  
122 using sodium arsenite ( $\text{NaAsO}_2$ ) (Panreac), dimethylarsinic acid ( $\text{C}_2\text{H}_7\text{AsO}_2$ ) (Sigma Aldrich),  
123 sodium methylarsonate ( $\text{CH}_3\text{AsO}(\text{ONa})_2 \cdot 6\text{H}_2\text{O}$ ) (Carlo Erba) and sodium arsenate  
124 ( $\text{Na}_2\text{HAsO}_4 \cdot 7\text{H}_2\text{O}$ ) (Panreac), respectively. Intermediate solutions of  $1 \text{ mg L}^{-1}$  were prepared  
125 daily and employed to obtain the calibration solutions with concentrations in the  $1\text{-}10 \text{ } \mu\text{g L}^{-1}$   
126 range. The mobile phase for liquid chromatography consisted of aqueous solutions of  $\text{K}_2\text{HPO}_4$

127 (Merck) and  $\text{KH}_2\text{PO}_4$  (Panreac). In addition, the generation of hydrides was carried out using  
128 aqueous solutions of HCl and  $\text{NaBH}_4$  (Panreac). All solutions were prepared with ultrapure  
129 Milli-Q (Millipore) water. NIST SRM 1568b (rice flour) and LGC7162 (strawberry leaves)  
130 reference materials were employed for validation of total arsenic determination by acid  
131 digestion. NIST SRM 1568b (with inorganic As, DMA and MMA certified content) was also  
132 employed for arsenic speciation analysis.

133

## 134 *2.2. Methodology for arsenic speciation and total arsenic analysis*

135 Speciation determination of the fruits was carried out by coupling liquid chromatography,  
136 hydride generation and atomic fluorescence spectrometry (HPLC-HG-AFS). For the separation  
137 of the arsenic species, an anion exchange column (Hamilton PRPX-100 150 mm x 4.1 mm, 10  
138  $\mu\text{m}$ ) was used. The mobile phase was 20  $\text{mmol L}^{-1}$   $\text{KH}_2\text{PO}_4/\text{K}_2\text{HPO}_4$  (pH 5.8, flow rate 0.8 mL  
139  $\text{min}^{-1}$ ), pumped with a Shimadzu PU-2080 Plus pump. The order of elution was As(III), DMA,  
140 MMA and As(V), and separation was accomplished in less than 10 min. After separation of the  
141 arsenical species, they were converted into the corresponding hydrides by adding solutions of  
142 4  $\text{mol L}^{-1}$  HCl and 1.5 % (w/v)  $\text{NaBH}_4$  (stabilized in 1% (w/v) NaOH), both at a flow rate of 1  
143  $\text{mL min}^{-1}$ . A PS Analytical Millennium Excalibur 10.055 detector equipped with a Photron  
144 Superlamp was employed for AFS detection of the arsines. Data were recorded using Clarity  
145 Chromatography Software 2.6.2.226 from DataApex. A chromatogram of the arsenic species  
146 obtained by HPLC-HG-AFS is shown in the Supplementary Material (see Fig. S1). The  
147 concentration range to obtain the linear calibration graphs for each individual arsenic species  
148 was 1-10  $\mu\text{g L}^{-1}$ . These instrumental settings have been previously reported for arsenic  
149 speciation studies in water and biological samples (Sánchez-Rodas, Geiszinger, Gómez-Ariza,  
150 & Krancesconi, 2002; Kohfahl, Sánchez-Rodas Navarro, Mendoza, Vadillo, & Giménez-

151 Forcada, 2016). The determination of the total arsenic content of the samples (roots, stems and  
152 fruits) was achieved after acid digestion and analysis with HG-AFS.

153

### 154 *2.3. Strawberry cultivation and addition of arsenic via root irrigation*

155 The strawberry plants were cultivated at an experimental plot of Adesva (Technological  
156 Center of the Agribusiness) located in Lepe (Huelva, Spain). Hydroponic cultivation of the  
157 plants was performed using commercially available coconut fibre as a substrate. Cultivation  
158 started in October, and the addition of inorganic arsenic irrigation was conducted via irrigation  
159 for one month (May), once the fruition had started.

160

161 Six experiments were performed on the addition of arsenic species to the strawberry  
162 cultivation via root irrigation (Table 1). They corresponded to the addition of either As(III) or  
163 As(V) to the irrigation water at concentrations of 10, 100 or 1000  $\mu\text{g L}^{-1}$ . For each experiment,  
164 two sacks of substrate were used, and each sack contained eight strawberry plants. An aliquot  
165 of 50 mL of each arsenic solution was added manually to each strawberry plant daily, using  
166 plastic syringes. A cultivation blank sample without arsenic addition was also considered.

167

### 168 *2.4. Sampling and sample treatment*

169 The sampling of the fruits was carried out once per week during May. A composite  
170 sample of ca. 60-70 g corresponding to 4 ripe strawberries was obtained from each of the six  
171 arsenic cultivation experiments and the blank sample. Strawberries were randomly picked from  
172 different plants. The amount of fruit collected every sampling day was between 420-490 g. The  
173 strawberries were frozen and lyophilized for ca. 96 hours (Telstar model Cryodos-80). The  
174 lyophilized strawberries were crushed and homogenized. A total of 24 arsenic enriched fruit  
175 composite samples were obtained during the four week sampling period for the 6 addition

176 cultivation experiments, plus 4 composite samples for the culture blank in the same period. At  
177 the end of the four week sampling period, plants were uprooted. Roots and stems were  
178 lyophilized, crushed and homogenized in a similar way as the fruits. Arsenic concentrations in  
179 the three parts of the plants were referred to the dry weight of the samples after lyophilization.

180

181 For extraction of the arsenic species from the fruits, 200 mg of lyophilized strawberries  
182 was weighed and 10 mL of Milli-Q water was added. The extraction was aided by heating in a  
183 microwave oven at 90 W for 4 min. The extracts were centrifuged at 5300 rpm for 5 min. The  
184 supernatant was submitted to a clean-up step using a C18 cartridge. The eluate was filtered  
185 with a 0.45  $\mu\text{m}$  PVDF syringe filter. Finally, the extracts were analyzed by HPLC-HG-AFS to  
186 determine the four arsenic species.

187

188 Total arsenic determination in fruits, stems and roots was performed after acid digestion.  
189 A portion of 100 mg of homogenized dry sample was weighed in a Teflon digestion vessel and  
190 mineralized with 2 mL of 65%  $\text{HNO}_3$  and 2 mL of  $\text{H}_2\text{O}_2$  at 180  $^\circ\text{C}$  by heating on a hot plate.  
191 The residue was dissolved in 2%  $\text{HNO}_3$  to a final volume of 10 mL. Subsequently, the total  
192 arsenic content was determined with HG-AFS.

193

194 Parametric statistical test (Student's t-test) was applied to the arsenic accumulation data  
195 in the strawberry plants, using software package STATISTICA 7.0 (Statsoft). An  $\alpha$ -value of  
196 0.05 was adopted as the critical level for statistical testing giving a 95% confidence level.

197

### 198 **3. Results and Discussion**

199 *3.1 Validation of analytical procedures for total arsenic and arsenic speciation.*

200 The total arsenic content of the strawberry samples was determined by acid digestion  
201 followed by HG-AFS analysis. The validation of this analytical procedure was conducted by  
202 analyzing reference materials. The total arsenic concentrations obtained for NIST 1568b rice  
203 flour and LGC7162 strawberry leaves ( $0.276 \pm 0.010 \text{ mg kg}^{-1}$  and  $0.25 \pm 0.02 \text{ mg kg}^{-1}$ ,  
204 respectively) did not significantly differ from the analytically derived certified values ( $0.285 \pm$   
205  $0.014$  and  $0.28 \pm 0.07 \text{ mg kg}^{-1}$ , respectively).

206

207 The arsenic speciation analysis of the samples was performed by aqueous extraction  
208 aided by microwave radiation and HPLC-HG-AFS analysis. The validation of the speciation  
209 procedure was performed by analyzing NIST 1568b rice flour. The concentration obtained for  
210 inorganic arsenic (sum of As(III) and As(V)) was  $0.091 \pm 0.001 \text{ mg kg}^{-1}$  and did not differ from  
211 the analytically derived certified value ( $0.092 \pm 0.001 \text{ mg kg}^{-1}$ ). In addition, the experimental  
212 values for DMA ( $0.181 \pm 0.001 \text{ mg kg}^{-1}$ ) and MMA ( $0.0118 \pm 0.0012 \text{ mg kg}^{-1}$ ) were similar to  
213 the certified concentrations ( $0.180 \pm 0.012 \text{ mg kg}^{-1}$  for DMA,  $0.0116 \pm 0.0035 \text{ mg kg}^{-1}$  for  
214 MMA). To confirm the effective extraction of the arsenic species from the strawberry fruits  
215 samples, an spiking experiment was conducted by duplicate additions of  $125 \mu\text{L}$  of a  $0.5 \text{ mg}$   
216  $\text{L}^{-1}$  solution of the four arsenic species (in methanol) to  $0.2 \text{ g}$  of a strawberry sample. After  
217 evaporation of the solvent, the samples were extracted with water, filtered, submitted to C18  
218 cartridge clean up and analyzed by HPLC-HG-AFS. The results indicated quantitative  
219 recoveries for all the arsenic species:  $95 \pm 5\%$  As(III),  $102 \pm 5\%$  DMA,  $99 \pm 3\%$  MMA and  $99 \pm 2\%$   
220 As(V).

221

### 222 *3.2. Accumulation of arsenic in roots, stems and fruits of strawberry plants*

223 The results of the cultivation experiments indicated that the accumulation of arsenic in  
224 the strawberry plants was dependent on the level of exposure and the particular arsenic species

225 added to the irrigation. The roots of the strawberry plants were found to have the highest  
226 amount of accumulated arsenic. The concentrations of arsenic in the roots increased  
227 progressively as the concentrations of As(III) and As(V) introduced to the irrigation increased  
228 (Table 1). In the cultivation experiments with low and medium arsenic concentrations (10 and  
229  $100 \mu\text{g L}^{-1}$ ), the accumulation trend was similar for both As(III) and As(V). In those cases, the  
230 arsenic concentration in the roots increased from 0.27 to 0.46  $\text{mg kg}^{-1}$  for As(III) exposure, and  
231 from 0.27 to 0.39  $\text{mg kg}^{-1}$  for As(V) exposure. However, when the plants were exposed to high  
232 arsenic concentrations ( $1000 \mu\text{g L}^{-1}$ ) the accumulation in the roots was ten-fold higher for  
233 As(III) than for As(V) ( $4.1 \text{ mg kg}^{-1}$  vs.  $0.44 \text{ mg kg}^{-1}$ ). The different accumulation rates of  
234 As(III) and As(V) can be attributed to the different mechanisms of uptake for each arsenic  
235 species in the plants. It has been reported that As(III) is efficiently taken up through the silicon  
236 transport pathway, probably by sharing nodulin26-like intrinsic proteins, whereas As(V) is  
237 analogous to phosphate and competes for the same transport in the roots (Nagajyoti, Lee, &  
238 Sreekanth, 2010; Zhao, McGrath, & Meharg, 2010; Wang, Peng, Tan, Ma, & Rathinsabathi,  
239 2015).

240

241 The accumulation in stems followed a similar trend as in roots (Table 1). Overall, the  
242 concentration of arsenic found in stems was significantly lower than in the roots (t-test,  
243  $p < 0.002$ ) for both As(III) and As(V) exposure, with the exception of As(V) at  $1000 \mu\text{g L}^{-1}$  (t-  
244 test,  $p > 0.7$ ). Once more, as expected, the highest accumulation was observed for  $1000 \mu\text{g L}^{-1}$   
245 arsenic exposure. The accumulation was again higher for As(III) ( $1.27 \text{ mg kg}^{-1}$ ) than for As(V)  
246 ( $0.43 \text{ mg kg}^{-1}$ ). The concentration ratio  $A_{\text{Sroots}}/A_{\text{Sstems}}$  was in most cases higher than 1.0,  
247 indicating a preferential accumulation in the roots rather than in the stems. When exposed to  
248 the highest arsenic concentration, the ratio  $A_{\text{Sroots}}/A_{\text{Sstems}}$  was 3.2 for As(III) exposure, whereas  
249 a lower ratio  $A_{\text{Sroots}}/A_{\text{Sstems}} = 1.0$  was obtained for As(V) exposure. The preferential

250 accumulation in the roots has also been found in other edible plants, such as broad bean plants  
251 (Sadee, Foulkes, & Hill, 2016). However, when other edible plants, such as carrots, are exposed  
252 to arsenic-rich water, the accumulation is higher in the leaves than the roots (Mayorga,  
253 Moyano, Anawar, & García-Sanchez, 2013).

254

255 The accumulation of arsenic in the fruits was lower than it was in the roots and stems. Of  
256 the three levels of arsenic exposure, accumulation in the fruits occurred only with the highest  
257 level of arsenic ( $1000 \mu\text{g L}^{-1}$ ) exposure. The results in Table 2 correspond to the total arsenic  
258 concentrations of the samples after acid digestion and analysis with HG-AFS. A progressive  
259 accumulation of arsenic was observed throughout the four weeks of the studied period for both  
260 species. As it happened in roots and stems, the accumulation was greater for As(III) than for  
261 As(V) exposure. This preferential accumulation of As(III) compared to As(V) was significantly  
262 higher after the 3<sup>rd</sup> and 4<sup>th</sup> weeks of exposure (t-test,  $p < 0.04$  and  $p < 0.03$ , respectively), but not  
263 during the 1<sup>st</sup> and the 2<sup>nd</sup> weeks (t-test,  $p > 0.7$  and  $p > 0.18$ , respectively). The concentration of  
264 arsenic in the fruits increased from  $0.06$  to  $0.30 \text{ mg kg}^{-1}$  when As(III) is applied, and from  $0.04$   
265 to  $0.22 \text{ mg kg}^{-1}$  for As(V).

266

267 The different accumulation rate of arsenic in fruits, stems and roots can be due to a  
268 detoxification mechanism, similarly as proposed for other plants. For carrots, it has been  
269 indicated that arsenate absorbed by roots is reduced to arsenite and transported to leaves via  
270 chelating agents (e.g. phytochelatin, GSH), and thus the toxicity of arsenic is reduced. This  
271 translocation generally can explain the amount of arsenic in different plant organs. (Mayorga,  
272 Moyano, Anawar, & García-Sanchez, 2013). The location of the As-complexes within plant  
273 tissue is not clear known, although it is known that these complexes are stable in vacuoles

274 under acidic conditions. The transporters for uptake in leaf cells are assumed to be similar as  
275 those in roots (Zhao, McGrath, & Meharg, 2010).

276

### 277 *3.3. Analysis of speciation of arsenic in the strawberry fruits*

278 Speciation analysis of arsenic for As(III), As(V), MMA and DMA was successfully  
279 carried out in the fruits after water extraction of the samples aided by microwave radiation and  
280 analysis by HPLC-HG-AFS. Speciation analysis in the fruits was only possible in the  
281 experiments corresponding to exposure to 1000  $\mu\text{g L}^{-1}$  of As(III) and As(V). No accumulation  
282 in the fruits was found for exposure at 10 or 100  $\mu\text{g L}^{-1}$  arsenic. The results of speciation  
283 analysis of the fruits of these two experiments during the four weeks of study are shown in  
284 Table 3. Speciation analysis was also attempted in the roots and stems, but no arsenic species  
285 were detected in the aqueous extracts. This difficulty in effectively extracting arsenic species  
286 with water or water/methanol mixtures from these parts of terrestrial plants been previously  
287 reported, reporting extraction efficiencies as low as 3% (Ruiz-Chancho, López-Sánchez,  
288 Schmeisser, Goessler, Francesconi, & Rubio, 2008).

289

290 In the experiment corresponding to the addition of 1000  $\mu\text{g L}^{-1}$  of As(III), no arsenic  
291 species were detected in the first week, as their concentrations were below 0.01  $\text{mg kg}^{-1}$ . As(III)  
292 was detected in the second week of study and its concentration increased until the fourth week  
293 (0.23  $\text{mg kg}^{-1}$ ). This increase was significant during the 2<sup>nd</sup> and 3<sup>rd</sup> week (t-Student,  $p < 0.03$ ),  
294 but not the 4<sup>th</sup> week (t-Student,  $p > 0.2$ ). As(III) was partially converted into DMA and MMA  
295 and oxidized to As(V) at the end of the study period. Fig. S2 in the Supplementary Materials  
296 shows a chromatogram obtained by HPLC-HG-AFS corresponding to the analysis of arsenic  
297 speciation in the fruit in the fourth week of the study. The presence of methylated arsenic  
298 species in plants is usually explained by a pathway involving S-adenosylmethyltransferase that

299 act on the methyl donor S-adenosyl-L-methionine (SAM) and As(III) as initial substrate (Zhao,  
300 McGrath, & Meharg, 2010). In addition, the methylation of inorganic arsenic to DMA in some  
301 plants (e.g., rice) can be the result of rhizosphere associated bacteria (Arao, Kawasaki, Baba,  
302 & Matsumoto, 2011).

303

304 With respect to the addition of 1000  $\mu\text{g L}^{-1}$  As(V) to the irrigation, the results showed a  
305 lower accumulation of inorganic arsenic as well as a lower transformation rate to methylated  
306 species. It was observed that As(V) was strongly reduced to As(III). This pattern has also been  
307 described in rice plants, indicating that arsenate is readily reduced to arsenite, which can later  
308 be detoxified by complexation with thio-rich peptides such as phytochelatin (Zhao, McGrath,  
309 & Meharg, 2010). The latter methylated species were detected in the fruit from the second  
310 week and increased until the fourth week ( $0.14 \text{ mg kg}^{-1}$ ). Arsenic (V) was only detected in the  
311 fourth week at a concentration of  $0.03 \text{ mg kg}^{-1}$ . In the case of methylated species, DMA was  
312 not detected throughout the study period while MMA was detected in the last two weeks of the  
313 experiment.

314

315 The speciation results indicated that the sum of methylated DMA and MMA represent  
316 21% of the total arsenic content for As(III) exposure, the remaining 79% corresponding to  
317 inorganic arsenic. The percentage of methylated species was found to be a 5% (corresponding  
318 to MMA) for As(V) exposure. Other studies performed with rice grains found similar  
319 percentages for rice cultivation in Asia (up to 78% inorganic arsenic), whereas crops from  
320 Europe and USA showed an increasing trend of DMA in relation to total arsenic (Wang, Peng,  
321 Tan, Ma, & Rathinsabathi, 2015). The methylation in the fruit was lower than in other plants,  
322 such as the edible bean, in which DMA (19%) and MMA (48%) are the dominant species  
323 (Sadee, Foulkes, & Hill, 2016).

324

325           The sum values of the inorganic species after four weeks of 1000  $\mu\text{g L}^{-1}$  exposure (Table  
326 3) are 0.25  $\text{mg kg}^{-1}$  and 0.17  $\text{mg kg}^{-1}$  (with respect to dry weight) for irrigation with As(III) and  
327 As(V), respectively. The water content of the fruits was 89%, calculated by weighing them  
328 before and after lyophilization. Considering this water content, the arsenic concentration in the  
329 fresh fruits corresponded to 27.5  $\mu\text{g As kg}^{-1}$  for irrigation with spiked As(III) and 11.2  $\mu\text{g As}$   
330  $\text{kg}^{-1}$  for irrigation with spiked As(V). These results indicated that the arsenic content of the  
331 fruits does not represent a health hazard. It would be necessary to consume ca. 6.0-8.8 kg of  
332 strawberries daily to reach the limit established by WHO for arsenic of 3.0  $\mu\text{g As kg}^{-1}$  body  
333 weight per day (WHO, 2011b), considering an average body weight of 55 kg. The presented  
334 results showed that even under a high exposure of the plants to As(III) or As(V), the inorganic  
335 arsenic content of the fruits may not represent a serious health risk in the case of human  
336 consumption.

337

### 338 **Conclusions**

339           This study provides information based on experimental data to better understand the  
340 accumulation and transformation of arsenic species in different parts of the strawberry plants  
341 (root, stem and fruits) cultivated on a hydroponic substrate. The cultivation experiments  
342 performed with arsenic exposure via root irrigation indicated a greater uptake for As(III) than  
343 for As(V). The accumulation depended on the concentration of arsenic species added to the  
344 water and the time of exposure. The maximum accumulation was higher for roots than for stems  
345 and fruits. The arsenic speciation analysis of the fruits revealed that inorganic arsenic underwent  
346 partial methylation. Additionally, a marked reduction of As(V) to As(III) was observed in the  
347 fruits when As(V) was applied to the plants. Overall, the inorganic arsenic accumulated in the  
348 fruits did not represent a health risk in the case of human consumption.

349

## 350 **Acknowledgements**

351 This work has been financially supported by project FQM 752 of the Andalusian  
352 Autonomous Government (Consejería de Economía, Innovación, Ciencia y Empleo).

353

## 354 **References**

355 Abedin, M. J., Cresser, M. S., Meharg, A. A., Feldmann, J., & Cotter-Howells, J. (2002).  
356 Arsenic accumulation and metabolism in rice (*Oryza sativa* L.). *Environmental Science and*  
357 *Technology*, *36*, 962-968.

358

359 Abdul, K. S. M., Jayasinghe, S. S., Chandana, E. P. S., Jayasumana, C., & De Silva, M. C. S.  
360 (2015). Arsenic and human health effects: a review. *Environmental Toxicology and*  
361 *Pharmacology*, *40*, 828-846.

362

363 Arao, T., Kawasaki, A., Baba, K., & Matsumoto, S. (2011). Effects of arsenic compound  
364 amendment on arsenic speciation in rice grain. *Environmental Science and Technology*, *45*,  
365 1291-1297.

366

367 Bergqvist, C., Herbert, R., Persson, I., & Greger, M. (2014). Plants influence on arsenic  
368 availability and speciation in the rhizosphere, roots and shoot of three different vegetables.  
369 *Environmental Pollution*, *184*, 540-546.

370

371 Cerveira, C., Pozebon, D., Poméu de Moraes, D., & Silva de Fraga J. C. (2015). Speciation of  
372 inorganic arsenic in rice using hydride generation atomic absorption spectrometry (HG-AAS).  
373 *Analytical Methods*, *7*, 4528-4534.

374

375 Chen, B., Corns, W.T., Stockwell, P.B., and Huang J.H. (2014). Accurate fast screening for  
376 total and inorganic arsenic in rice grains using hydride generation atomic fluorescence  
377 spectrometry (HG-AFS). *Analytical Methods*, 6, 7554-7558

378

379 Dos Santos, G. M., Pozebon, D., Cerveira, C., & de Moraes, D. P. (2017). Inorganic arsenic  
380 speciation in rice products using selective hydride generation and atomic absorption  
381 spectrometry (AAS). *Microchemical Journal*, 133, 265-271.

382

383 European Food Safety Agency (EFSA) (2009). Scientific opinion on arsenic in food. EFSA  
384 Panel on Contaminants in the Food Chain (CONTAM). *EFSA Journal* 7(10):1351

385

386 Farías, S. S., Londonio, A., Quintero, C., Befani, R., Soro, M., & Smichowski, P. (2015). On-  
387 line speciation and quantification of four arsenical species in rice samples collected in  
388 Argentina using HPLC-HG-AFS coupling. *Microchemical Journal*, 120, 34-39.

389

390 Food and Agricultural Organization (FAO) of the United Nations. Faostat. (2019).  
391 <http://www.fao.org/faostat/en/#data/QC> /Accessed 20 March 2019.

392

393 Gupta, A., Nayak, S., Agarwal, R., Dobhal, D.P., Uniyal, P., & Singh, R. (2012). Singh Arsenic  
394 speciation analysis and remediation techniques in drinking water. *Desalination and Water  
395 Treatment*, 40, 231-243.

396

397 International Agency for Research on Cancer (IARC) (2012). Arsenic and arsenic compounds.  
398 In WHO (Ed.), *Arsenic, metals, fibres and dust. Volume 100C. A review on human carcinogens*  
399 (pp. 41-94). Lyon: International Agency for Research on Cancer.

400

401 Jedynek, L., & Kowalska, J. (2011). Stability of arsenic species in hydroponic media and its  
402 influence on arsenic uptake and distribution in White mustard (*Sinapis alba* L.). *Microchemical*  
403 *Journal*, 98, 163-169.

404

405 Kohfahl, C., Sánchez-Rodas Navarro, D., Mendoza, J. A., Vadillo, I., & Giménez-Forcada, E.  
406 (2016). Algae metabolism and organic carbon in sediments determining arsenic mobilisation in  
407 ground- and surface water. A field study in Doñana National Park, Spain. *Science of the Total*  
408 *Environment*, 544, 874-882.

409

410 Kuehnelt, D., Lintschinger, J., & Goessler, W. (2000). Arsenic compounds in terrestrial  
411 organisms. IV. Green plants and lichens from an old arsenic smelter site in Austria. *Applied*  
412 *Organometallic Chemistry*, 14, 411-420.

413

414 Kumarathilaka, P., Seneweera, S., Meharg, A., & Bundschuh, J. (2018). Arsenic speciation  
415 dynamics in paddy rice soil-water environment: sources, physic-chemical, and biological  
416 factors – A review. *Water Research*, 140, 403-414.

417

418 Lai, G., Chen, G., & Chen, T. (2016). Speciation of AsIII and AsV in fruit juices by dispersive  
419 liquid microextraction and hydride generation-atomic fluorescence spectrometry. *Food*  
420 *Chemistry*, 190, 158-163.

421

422 Ma, L., Yang, Z., Kong, Q., & Wang, L. (2017). Extraction and determination of arsenic species  
423 in leafy vegetables: Method development and application. *Food Chemistry*, 217, 524- 530.

424  
425 MAPAMA (Ministerio de Agricultura, Pesca, Alimentación y Medio Ambiente). Anuario  
426 (2011). [http://www.mapama.gob.es/estadistica/pags/anuario/2011/AE\\_2011\\_13.pdf](http://www.mapama.gob.es/estadistica/pags/anuario/2011/AE_2011_13.pdf) /  
427 Accessed 20 March 2019.

428  
429 Mattusch, J., Wennrich, R., Schmidt, A., Reisser, C., & Fresenius, W. (2000). Determination  
430 of arsenic species in water, soils and plants. *Fresenius Journal of Analytical Chemistry*, 366,  
431 200-203.

432  
433 Mayorga, P., Moyano, A., Anawar, H. M., & García-Sánchez, A. (2013). Uptake and  
434 accumulation of arsenic in different organs of carrot irrigated As-rich water. *Clean Soil Air*  
435 *Water*, 41(6), 587-592.

436  
437 Munoz, O., Díaz, P. B., Leyton, I., Nunez, N., Devesa Suñer, M.A., Velez, D., & et al. (2002).  
438 Vegetables Collected in the Cultivated Andean Area of Northern Chile: Total and Inorganic  
439 Arsenic Contents in Raw Vegetables. *Journal of Agricultural and Food Chemistry*, 50, 642–  
440 647.

441  
442 Nagajyoti, P.C., Lee, K.D., & Sreekanth, T.V.M. (2010). Heavy metals, occurrence and toxicity  
443 for plants: a review. *Environmental Chemistry Letters*, 8, 199-216.

444

445 Petursdottir, A.H., Sloth, J.J., & Feldmann, J. (2015). Introduction of regulations for arsenic in  
446 feed and food with emphasis on inorganic arsenic, and implications for analytical chemistry.  
447 *Analytical and Bioanalytical Chemistry*, 407, 8385-8396.

448

449 Ruiz-Chancho, M.J., López-Sánchez, J.F., Schmeisser, E., Goessler, W., Francesconi, K.A., &  
450 Rubio, R. (2008). Arsenic speciation in plants growing in arsenic-contaminated sites.  
451 *Chemosphere*, 71, 1522-1530.

452

453 Sadee, B. A., Foulkes, M. E., & Hill, S. J. (2016). A study of arsenic speciation in soil, irrigation  
454 water and plant tissue: A case study of the broad bean plant, *Vicia faba*. *Food Chemistry*, 210,  
455 362-370.

456

457 Sánchez-Rodas, D., Geiszinger, A., Gómez-Ariza, J. L., & Francesconi, K. A. (2002).  
458 Determination of an arsenosugar in oyster extracts by liquid chromatography-electrospray mass  
459 spectrometry and liquid chromatography-ultraviolet photo-oxidation-hydride generation  
460 atomic fluorescence spectrometry. *Analyst*, 127, 60-65.

461

462 Shahid, M., Rafi, M., Niazi, N. K., Dumat, C., Shanshad, S., Khalid, S., et al. (2017). Arsenic  
463 accumulation and physiological attributed of spinach in the presence of amendments:  
464 implication to reduce health risk. *Environmental Science Pollution Research*, 24, 16097-16106.

465

466 Sigrist, M., Hilbe, N., Brusa, L., Campagnoli, D., & Beldomenico, H. (2016). Total arsenic in  
467 selected food samples from Argentina: Estimation of their contribution to inorganic arsenic  
468 dietary intake. *Food Chemistry*, 210, 96- 101.

469

470 Sirot, V., Guérin, T., Volatier, J. L., & Leblanc, J. C. (2009). Dietary exposure and biomarkers  
471 of arsenic in consumers of fish and shellfish from France. *Science of the Total Environment*,  
472 407, 1875-1885.

473

474 Sons, S. H., Lee, W. B., Kim, D., Lee, Y., & Nam, S.H. (2019). An alternative analytical method  
475 for determining arsenic species in rice by using ion chromatography and inductively coupled  
476 plasma-mass spectrometry. *Food Chemistry*, 279, 353-358.

477

478 Taylor, V., Goodale, B., Raab, A., Schwerdtle, T., Reimer, K., Conklin, S., Karagas, M.R., &  
479 Francesconi K.A. (2017). Human exposure to organic arsenic species from seafood. *Science of*  
480 *the Total Environment*, 580, 266-282.

481

482 Ure, A. M., & Davidson, C. M. (2002). Chapter 10. Chemical speciation in soils and related  
483 materials by selective chemical extraction. In Ure, A. M., & Davidson, C. M. (Eds.), *Chemical*  
484 *Speciation in the Environment* (2<sup>nd</sup> ed.) (pp 265–300). Bodmin: Blackwell Science.

485

486 Upadhyay, M. K., Shukla, A., Yadav, P., & Srivastava, S. (2019). A review of arsenic in crops,  
487 vegetables, animals and food products. *Food Chemistry*, 276, 608-618.

488

489 Urgast, S., Adams, G. C., Raab, A., & Feldmann, J. (2010). Arsenic concentration and  
490 speciation of the marine hyperaccumulator whel *Buccinum undatum* collected in coastal waters  
491 of Northern Britain. *Journal Environmental Monitoring*, 12(5), 1126-32.

492

493 Wang, X., Peng, B., Tan, C., Ma, L., & Rathinsabathi, B. (2015). Recent advances in arsenic  
494 bioavailability, transport and speciation in rice. *Environmental Science Pollution Research*,  
495 22(8), 5742-50.  
496

497 World Health Organization (WHO) (2011a). Chapter 8. Chemical aspects. In: *Guidelines for*  
498 *drinking-water quality* (4<sup>th</sup> ed.) (pp. 155-201). Malta: World Health Organization.  
499

500 World Health Organization (WHO) (2011b). Seventy-second report of the Joint FAO/WHO  
501 Expert Committee on food additives. Evaluation of certain contaminants in food. Tech. Rep.  
502 Ser., 959 pp. 1-105.  
503

504 Zhao, F. J., McGrath, S. P., & Meharg, A. A. (2010). Arsenic as a food chain contaminant:  
505 Mechanisms of plant uptake and metabolism and mitigation strategies. *Annual Review of Plant*  
506 *Biology*, 61, 535-559.  
507

508 Zhao, D., Li, H. B., Xu, J. Y., Luo, J., & Ma, L. Q. (2015). Arsenic extraction and speciation in  
509 plants: Method comparison and development. *Science of the Total Environment*, 523, 138–145.